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1 **Regional and global perspectives of honey as a record of lead in the environment**

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4 **44 ABSTRACT**

5 45 Honey from *Apis mellifera* is a useful and inexpensive biomonitor for mapping metal
6 46 distributions in urban centers. The sampling resolution of a biomonitoring survey (e.g., city
7 47 versus global scale) determines which geochemical processes are reflected in the results. This
8 48 study presents Pb isotopic compositions and metal concentrations in honey from around the
9 49 world, sampled at varying resolutions: honey from Canada (n = 21), the United States (n = 111),
10 50 Belgium (n = 25), and New Zealand (n = 10), with additional samples from Afghanistan, Brazil,
11 51 Cuba, Germany, Liberia, Taiwan, and Turkey. Honey was sampled at high resolution in two
12 52 uniquely different land-use settings (New York Metro Area and the Hawaiian island of Kaua'i),
13 53 at regional-scale resolution in eastern North America (including the Great Lakes region), and Pb
14 54 isotopic compositions of all samples were compared on a global scale. At high sampling
15 55 resolution, metal concentrations in honey reveal spatially significant concentration gradients: in
16 56 New York City, metals associated with human activity and city infrastructure (e.g., Pb, Sb, Ti, V)
17 57 are more concentrated in honey collected within the city compared to honey from upstate New
18 58 York, and metal concentrations in honey from Kaua'i suggest polluting effects of nearby
19 59 agricultural operations. At lower resolution (regional and global scales), lead isotopic
20 60 compositions of honey are more useful than metal concentrations in revealing large-scale Pb
21 61 processes (e.g., the enduring legacy of global leaded gasoline use throughout the twentieth
22 62 century) and the continental origin of the honey. Lead isotopic compositions of honey collected
23 63 from N. America (especially from the eastern USA) are more radiogenic ($^{206}\text{Pb}/^{207}\text{Pb}$: 1.132-
24 64 1.253, $^{208}\text{Pb}/^{206}\text{Pb}$: 2.001-2.129) compared to European honey, and honey from New Zealand,
25 65 which has the least radiogenic isotopic compositions measured in this study ($^{206}\text{Pb}/^{207}\text{Pb}$: 1.077-
26 66 1.160, $^{208}\text{Pb}/^{206}\text{Pb}$: 2.090-2.187). Thus, biomonitoring using honey at different resolutions
27 67 reflects differing processes and, to some extent, a honey *terroir* defined by the Pb isotopic
28 68 composition. The data presented here provide important (and current) global context for future
29 69 studies that utilize Pb isotopes in honey. Moreover, this study exhibits community science in
30 70 action, as most of the honey was collected by collaborators around the world, working directly
31 71 with local apiarists and hobby beekeepers.
32 72

33 **73 KEYWORDS**

34 74 honey, lead isotopes, New York City, Kaua'i, global lead array
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1. INTRODUCTION

Since the Industrial Revolution, natural emissions of certain metals and metalloids (e.g., Pb, V, Sb, Hg, Ni, Cd, Zn) have been dwarfed by anthropogenic emissions (Han et al., 2002; Nriagu, 1989; Rauch and Pacyna, 2009). Of ongoing concern is lead (Pb), a neurotoxic metal and known teratogen, which is difficult (and costly) to remediate once deposited in the environment (Landrigan et al., 2018; Mielke et al., 2011; O'Connor et al., 2020). Natural sources of Pb in the environment include volcanic emissions, sea spray, and weathering and transport of local rocks, soils, and sediments. Anthropogenic inputs include emissions from mining operations, smelting, fuel combustion (low-grade coal, fuel oil, gasoline), and battery manufacturing and recycling. The most notable modern anthropogenic inputs of Pb to the environment are the use of leaded gasoline throughout most of the twentieth century and leaded paints resulting in the contamination of indoor dust and residential topsoil (affecting gardens and playgrounds) (Filippelli and Laidlaw, 2010; Mielke, 2018; Mielke and Reagan, 1998; O'Connor et al., 2018). The chemical 'legacy' of leaded gasolines and paints govern Pb-focused environmental studies to this day, decades after termination of their widespread use, including soil remediation efforts and public health concerns relating to elevated blood lead levels, especially in children (Filippelli et al., 2015; Zahran et al., 2013).

The results of modern lead use are apparent on local and global scales. From point sources, many of which release coarse particulates ($> 10 \mu\text{m}$), the extent of environmental enrichment for Pb is local or regional since coarse particulates have shorter atmospheric residence time, minimizing transport and causing an exponential decrease in pollutant concentration in topsoils as a function of distance from the source (Dong et al., 2020; Masri et al., 2015; Reimann et al., 2011). For example, degradation of exterior paints, resuspension of road dust or contaminated soil, or mechanical wear from metal refining processes all tend to cause local Pb pollution (Laidlaw and Filippelli, 2008; Weiss et al., 2006). Conversely, fine particulates ($< 2.5 \mu\text{m}$) have longer residence times in the atmosphere before settling/deposition, so they are more susceptible to transport than coarse particulates (Wilson et al., 2005). The consequences of leaded gasoline use are evident on local, continental, hemispheric, and global scales because combustion of leaded gasoline produces both coarse particulates that are deposited along roadways, especially in areas of high traffic (i.e., cities),

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162 and fine particulates, approximately 35 % of which fall into the ‘ultrafine’ size fraction (< 0.1
163 μm) (Mielke et al., 2011) that are subject to long-range (i.e., transcontinental) transport. In
164 general, high-resolution geochemical mapping reveals metal distribution at smaller scales,
165 highlighting the impact of local point or diffuse sources, which tend to be more anthropogenic
166 (e.g., Pb from an urban center, a smelter, or a mining operation) (Demetriades et al., 2010;
167 Locutura and Bel-lan, 2011; Mielke, 1994). Lower-resolution studies reveal continental and
168 global-scale processes, most of which are natural (i.e., crustal geochemistry, glaciation
169 boundaries, large-scale climate trends) (Smith et al., 2018), but also reflect some anthropogenic
170 activity, of which the global use of leaded gasoline is a key component.

171 The extent and history of Pb pollution is recorded in natural environmental archives and
172 the analysis of Pb stable isotopes (^{208}Pb , ^{207}Pb , ^{206}Pb , ^{204}Pb) in these archives aids researchers in
173 differentiating Pb sources (natural, anthropogenic, and legacy anthropogenic) (e.g., Komárek et
174 al., 2008) on various scales. Environmental proxies for Pb with local or regional spatiotemporal
175 constraints include lichens and mosses (Simonetti et al., 2003), fish and shellfish (Li et al., 2020;
176 Shiel et al., 2012), and tree rings (Marcantonio et al., 1998; Novak et al., 2010; Patrick and
177 Farmer, 2006). Sediment, peat, ice, and snow cores provide anywhere from continental to
178 global-scale records of Pb pollution (e.g., Hong et al., 1994; McConnell et al., 2018; Shotyky et al.,
179 1998; Weiss et al., 1999). Agricultural and consumer products (e.g., wine, vinegar, tobacco)
180 have Pb isotopic compositions that reflect their local environment of origin, thus serving as an
181 ‘archive’ of the Pb chemistry where the product originated (Epova et al., 2020; Guo et al., 2015;
182 Kristensen et al., 2016; Medina et al., 2000; Ndung’u et al., 2011).

183 Honey, and European honeybees (*Apis mellifera*), are effective biomonitors of chemicals
184 in the environment. Honey and bees have been used for monitoring pesticide residues on or
185 near farming operations (e.g., Codling et al., 2016; de Oliveira et al., 2016; Porrini et al., 2003)
186 and for determining metal distributions (including Pb) near point sources of metal pollution
187 (Smith et al., 2019; Zhou et al., 2018b). Since bees forage within a few kilometers of their hive
188 (Eckert, 1933) and passively collect dust and particulates while they forage for pollen and
189 nectar, the bees and their products (honey, bee pollen, wax) develop a chemical composition of
190 spatial significance, that reflects their foraging environment (Negri et al., 2015; Smith and Weis,

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4 191 2020; Taylor, 2019). City-scale investigations (tens of kilometers) of metal distributions using
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6 192 honey and hive products are numerous (e.g., Bromenshenk et al., 1985; Conti and Botrè, 2001;
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8 193 Giglio et al., 2017; Skorbiłowicz et al., 2018; Van der Steen et al., 2015; Zarić et al., 2016).
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10 194 However, source apportionment investigations using Pb isotopic compositions in honey (and
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12 195 other hive products) are still somewhat rare, and the existing studies are limited to city-scale
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14 196 and point-source investigations (Smith et al., 2019; Smith and Weis, 2020; Zhou et al., 2018b).
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16 197 Published metal concentration data compiled for honey from around the world is typically
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18 198 reported for food science and food quality purposes (e.g. Solayman et al. 2016), rather than for
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20 199 geochemical surveys, or verification of honey authenticity, which are sometimes used in
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22 200 conjunction with carbon stable isotopes or strontium isotopes (Baroni et al., 2015; Batista et al.,
23
24 201 2012; Zhou et al., 2018c). Missing from the literature is a global Pb isotopic database of honey.
25
26 202 For honey to be a viable biomonitor of Pb in the future, it is essential to provide a current,
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28 203 global context for existing and future city-scale studies.

29
30 204 In this study, we present applications of honey as a biomonitor for Pb in two disparate
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32 205 regions: the very populated and industrialized eastern North America (n = 58, featuring samples
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34 206 from the New York Metropolitan Area and the Great Lakes region), and the very remote
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36 207 Hawaiian Islands (n = 52). We then examine honey as a biomonitor for Pb on a global scale,
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38 208 using a suite of honey samples collected around the world (n = 181). Using these data, we
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40 209 demonstrate how honey fits within the modern global Pb array, as defined by previous regional
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42 210 and city-scale honey studies as well as anthropogenic and natural Pb isotopic compositions
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44 211 measured in consumer products and aerosols. The database for Pb isotopes in honey presented
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46 212 here validates a ‘community science’ approach, as many of the samples were sourced from
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48 213 collaborators and hobby apiarists around the world.

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50 51 215 **2. MATERIALS AND METHODS**

52 53 216 **2.1. Field and sampling information**

54 55 217 *2.1.1. Field methods*

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57 218 Honey samples were collected by community scientists who volunteered honey from their
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59 219 backyard apiaries or were purchased (as commercially available honey) in the country of origin

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220 (i.e., while traveling) or in Canada and the United States (USA) as imported products.

221 Community scientists used clean sampling vials, sampling instructions, and field worksheets for
222 consistent metadata collection, after Smith and Weis (2020).

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224 *2.1.2. Sampling sites*

225 Eastern North America (including the Great Lakes region) hosts the majority of the continent's
226 population and is rich in large urban centers, which feature concentrated areas of industrial
227 activity (U.S. EPA, 2017; Canada NPRI, 2017). The New York Metro Area includes the five
228 boroughs of New York City (NYC) and extends into parts of New Jersey (NJ), Pennsylvania (PA),
229 and Connecticut (CT) and is home to more than 23 million people (U.S. Census Bureau, 2016). A
230 community science effort, notably led by a high school student, provided honey samples from
231 NY Metro Area (n = 29) and samples from upstate NY, PA, and CT (n = 14), collected during
232 summer 2019.

233 The Hawaiian Islands offer a stark contrast to eastern North America. With a total
234 population of around 1.4 million, there are few urban and industrial (i.e., manufacturing)
235 centers on the Hawaiian Islands; the main industries are agriculture, tourism, and military
236 activities. The Hawaiian Islands are very remote (a volcanic island chain, located in the central
237 Pacific Ocean), and each island is small (< 10,500 km²), with a compact watershed and wet,
238 tropical climate. These factors, and year-round crop production, make the Hawaiian Islands a
239 unique setting to investigate honey as a biomonitor. Honey was sampled on the Island of Kaua'i
240 (31 samples from 22 unique sites) through community sampling efforts between 2013-2016
241 (Berg et al., 2018). These Kaua'i samples were collected directly from feral hives, managed
242 hives, or as commercial honey from local producers. For context, additional commercial honey
243 samples were purchased on the Islands of O'ahu, Maui, and Hawai'i (Big Island) in 2017-2018.
244 Other significant community sampling efforts for this study occurred in and around Brussels,
245 Belgium; Madison, Wisconsin (WI, USA); and in New Zealand (from North and South Island).

246 Mississippi Valley Type (MVT) galena samples (n = 4) from Lafayette County, WI, part of
247 the Upper Mississippi Valley Mining District, were provided by the WI Geological and Natural
248 History Survey. While not used in tetraethyl lead (TEL) production to the same extent as the
249 southeast (SE MO) MVT ores, these samples provide important geochemical context for Upper

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4 250 Mississippi Valley Pb ores used throughout the Great Lakes Megalopolis Manufacturing Belt
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6 251 during and after the Industrial Revolution.
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10 253 **2.2. Sample preparation and analysis**

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12 254 *2.2.1. Honey*

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14 255 Samples were prepared and analyzed in clean laboratories at the Pacific Centre for Isotopic and
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16 256 Geochemical Research (PCIGR) at the University of British Columbia (UBC), except for the honey
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18 257 and galena collected in Wisconsin (described in section 2.2.2). All acids reagents used at PCIGR
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20 258 were purchased as highest-purity grade available or were distilled in-house, and only ultrapure
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22 259 H₂O was used (18.2 MΩ·cm). Honey samples were digested and analyzed using the methods
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24 260 described by Smith et al. (2019): microwave digested in concentrated HNO₃, then evaporated
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26 261 and reconstituted in 2 % HNO₃ + 10 ng/g indium (In, added as an internal standard for metals
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28 262 analysis by inductively-coupled plasma mass spectrometry, ICPMS). Each honey digestion batch
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30 263 also included the standard reference material (SRM) NIST 1568b (Rice Flour, National Institute
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32 264 of Standards and Technology, Gaithersburg, MD, USA).

33 265 Metal concentrations (Mg, Al, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, Ga, As, Rb, Sr, Zr, Mo, Cd,
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35 266 Sn, Sb, Ba, Pb) were determined by ICPMS (Agilent 7700x, Agilent Technologies, Santa Clara, CA,
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37 267 USA). These elements include major and trace elements that originate from lithogenic and
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39 268 anthropogenic sources (or both), are measurable in honey, and have been useful in past studies
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41 269 for city-scale metal distribution surveys using honey (Smith et al., 2019; Smith and Weis, 2020).
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43 270 Lead isotopic compositions were measured in the same digests by high resolution (HR-) ICPMS
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45 271 (Nu AttoM, Nu Instruments Ltd., Wrexham, UK). In-house multicollector (MC-) ICPMS
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47 272 verification of NIST 1568b was previously presented by Smith and Weis (2020).
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51 274 *2.2.2. Wisconsin honey and galena*

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53 275 The Wisconsin honey (n = 7) and galena (n = 4) samples were prepared and analyzed at the
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55 276 Trace Element Clean Lab (TECL) facility at the Wisconsin State Laboratory of Hygiene, University
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57 277 of Wisconsin-Madison. All reagents were of Optima-grade, purchased from Thermo Fisher
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59 278 Scientific (Waltham, MA, USA). Honey samples were weighed into pre-cleaned 30 mL Teflon
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4 279 beakers (Savillex Corporation, MN, USA) and digested using 10 mL of 16 M HNO₃. Samples were
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6 280 dried, then re-digested in 9.9 mL 16 M HNO₃ + 0.1 mL 30% H₂O₂. One mL of this digest was
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8 281 removed and diluted in 2% HNO₃ for metals analysis using an Element2 HR-ICPMS (Thermo
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10 282 Fisher Scientific). The remaining solution was dried and reconstituted in 5 mL of 1 M HBr.
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12 283 Samples were loaded onto BioRad PolyPrep columns (BioRad Laboratories, Inc., CA, USA)
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14 284 containing 1 mL of pre-cleaned AG1-X8 anion exchange resin (BioRad) preconditioned in 1 M
15
16 285 HBr. Matrix elements were removed using washes of 1 M HBr, and the Pb fraction was
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18 286 collected using ultrapure H₂O and 1 M HNO₃. Galena samples were dissolved using 16 M HNO₃
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20 287 and were not purified prior to analysis. Final samples were diluted in 2 % HNO₃ and spiked with
21
22 288 Tl to correct for mass bias. Pb and Tl isotopes were analyzed using a Neptune Plus MC-ICPMS
23
24 289 (Thermo Fisher Scientific) in static mode using 6 faraday collectors, with an additional collector
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26 290 used to monitor for Hg interferences. Honey sample solutions were introduced into the plasma
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28 291 using an Aridus 3 desolvating nebulizer (Teledyne CETAC Technologies, NE, USA) and the Jet
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30 292 sampler and X-skimmer cone configuration. Galena sample solutions were introduced into the
31
32 293 plasma using the standard cyclonic spray chamber with the standard sampler cone and X-
33
34 294 skimmer cone configuration. Lead isotopic standard NIST 981 was analyzed periodically, and
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36 295 final isotope ratios are reported relative to NIST 981 values determined by Galer and
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38 296 Abouchami (1998). NIST 1568b was included for quality control.

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41 298 **2.4. Calculations and statistical analysis**

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43 299 All statistical calculations were made using R, v. 4.0.2 (R Core Team, 2020). Honey data from
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45 300 different global regions (e.g., Fig. 1) were compared using a Kruskal-Wallis rank sum test
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47 301 (Kruskal and Wallis, 1952). If the results were significant, indicated by an adjusted *p*-value (*p*_{adj.})
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49 302 < 0.05, then a post hoc Dunn's multiple comparison test (Dunn, 1964) was used to identify
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51 303 significantly different subgroups.

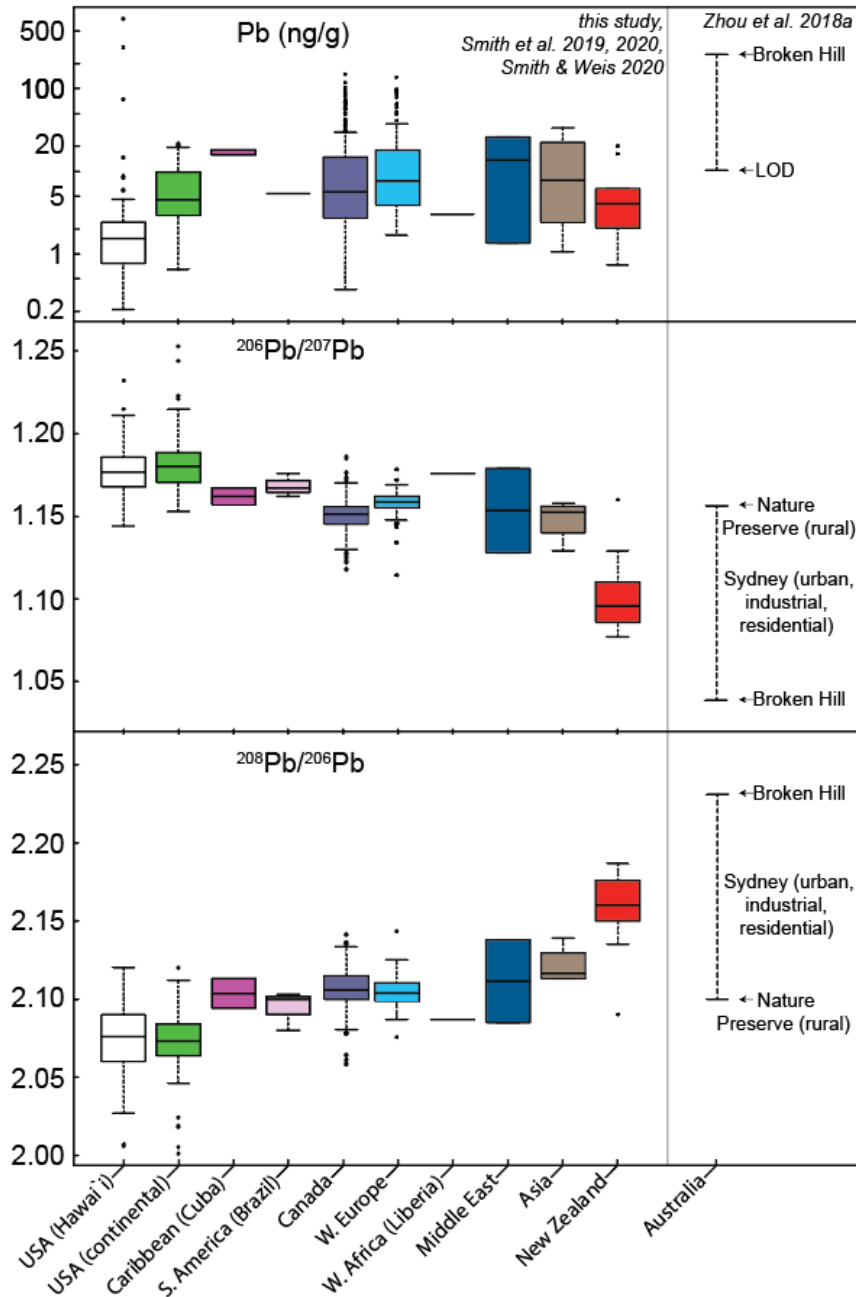


Figure 1. Box plots comparing Pb concentrations and isotopic compositions for honey, by region. The plotted data include this study (Table S1, S2), and other honey data from France (included in the 'W. Europe' region, Smith et al., 2020) and previous honey data reported for Canada (Smith et al., 2019; Smith and Weis, 2020). For comparison, Australian honey data from Zhou et al. (2018b) is also included; collected from Sydney (urban, industrial, and residential area), a nature preserve (rural), and from near the Broken Hill mine (LOD = limit of detection reported by Zhou et al., 2018b). Note that there is a logarithmic scale for the Pb concentration panel (top panel), only. A summary for pairwise statistical comparison, by region (for all regions with ≥ 10 measurements), for metal concentrations and Pb isotopes is included in the Supplementary Material (Table S3).

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315 **3. RESULTS**

316 **3.1. Blanks and standard reference materials**

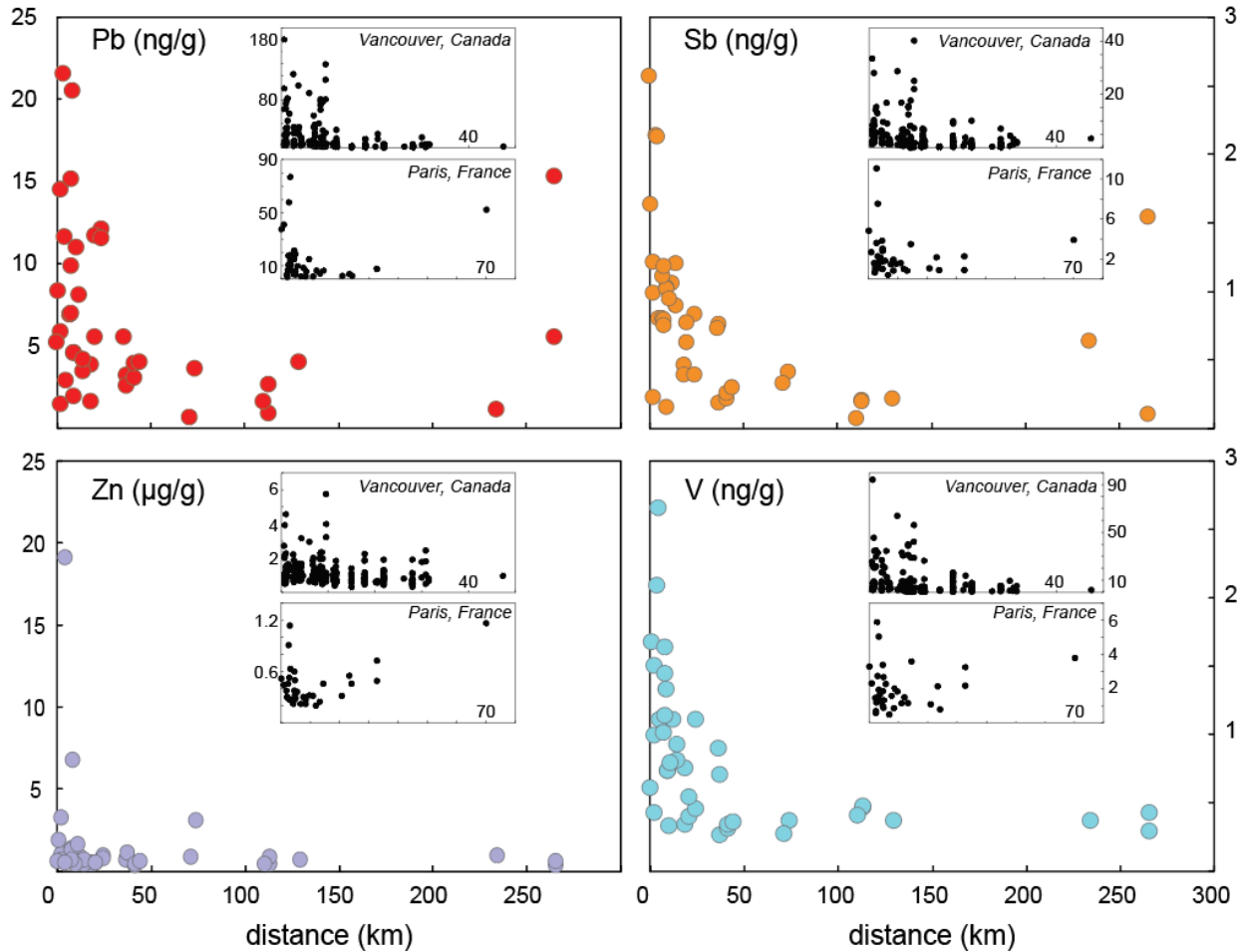
317 All metal concentrations, Pb isotope data, and summary for statistical tests are reported in the
318 Supplementary Material (Tables S1-S3). Trace element results for microwave digestion blanks
319 (honey digestion method) are reported in previous studies (Smith et al. 2019, Smith and Weis
320 2020, with an average of 36 pg total Pb). Refer to Smith and Weis (2020) for in-house HR-ICPMS
321 and MC-ICPMS results summaries for NIST 1568b (dataset doi:10.5683/SP2/Y9MKTN). An
322 aliquot of the same NIST 1568b used at the PCIGR (UBC) was analyzed at the TECL (UW) facility
323 (n = 2). Average MC-ICPMS results between the two facilities are in good agreement, with
324 relative mean differences of 3.84 per mille (‰), 0.19 ‰, and 2.75 ‰ for $^{208}\text{Pb}/^{204}\text{Pb}$,
325 $^{207}\text{Pb}/^{204}\text{Pb}$, and $^{206}\text{Pb}/^{204}\text{Pb}$, respectively. It is worthwhile to note that analytical results of
326 aliquots from a newer bottle of NIST 1568b (purchased in 2020) were not in good agreement
327 with previous results (from either facility), indicating that this SRM does not have a
328 homogenous Pb isotopic composition. This supports previous claims (e.g., Pohl et al., 2017;
329 Smith and Weis, 2020) that a suitable (well characterized and widely available) SRM would be
330 very useful for studies using hive products for metal biomonitoring purposes.

331
332 **3.2. Metal concentrations**

333 Metal concentrations in all honey analyzed in this study are comparable to ranges reported in
334 honey from all over the world, compiled by Solayman et al. (2016), and do not indicate any
335 concern for consumer health. Our results are also generally comparable to more recent
336 analyses of global honey by Zhou et al. (2018c), with some honey from this study exceeding
337 their reported ranges: Kaua’i, HI (Al, Ba, Fe), Metro New York (Fe), Brazil (Cu, Mg), Belgium (Rb),
338 and New Zealand (Rb). While the Pb concentration range for Pb in this study is large (0.2 to 709
339 ng/g), most of the honey samples (> 98 %, 179 out of 181) have Pb contents below 135 ng/g.
340 The suite of Hawaiian honey samples is unique in that it includes honey with the lowest (0.2
341 ng/g, Hawai’i Is.) and highest (709 ng/g, Kaua’i) concentrations of Pb measured in this study.
342 The two highest Pb concentrations measured in Kaua’i (709 and 317 ng/g Pb) exceed
343 concentrations previously measured in honey from urban centers: Vancouver, BC; Broken Hill,

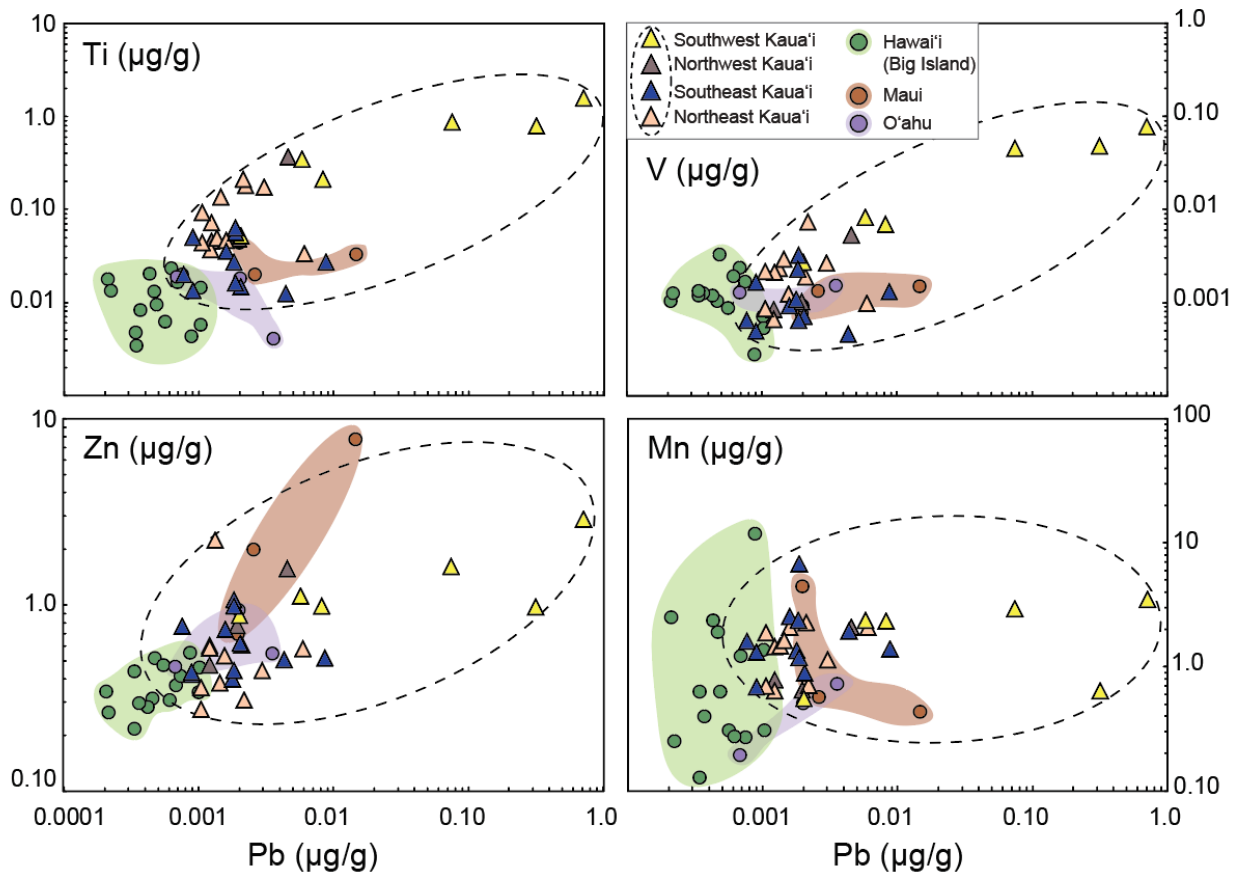
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344 Australia; Paris, France (Smith et al., 2020; Smith and Weis, 2020; Zhou et al., 2018b), but are
345 lower than Pb concentrations reported in honey in other studies from Europe, North Africa,
346 Malaysia, and the USA (Solayman et al., 2016 and references therein).



348 **Figure 2.** Metal concentrations in honey from the NY Metro Area as a function of radial distance
349 (km) from the center of Bryant Park, located in Midtown Manhattan. That site was selected to
350 demonstrate the effect on certain metal concentrations in honey as a function of distance from
351 the city center. Note that units for zinc (Zn) are µg/g, while others are ng/g. Concentration
352 maps for these metals in NY Metro honey are in the Supplementary Material (Fig. S1). For figure
353 clarity, error bars are not included (errors are reported in Table S1). For comparison, insets
354 show concentration plots for these metals in honey from two other cities: Metro Vancouver,
355 Canada (data compiled from Smith et al., 2019 and Smith and Weis, 2020) and Paris, France
356 (data from Smith et al., 2020). Distances are measured from the Port of Vancouver (after Smith
357 et al. 2019) in Metro Vancouver and from the Notre Dame cathedral in Paris. Insets have the
358 same units as their corresponding main plot for both axes.

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 4 360 For all metal concentrations, there is no systematic, geographic variation on a regional
 5 or global scale, (Table S3, Fig. 1), but in areas where sampling density was high (high sampling
 6 361 resolution), the local geospatial variations in metal concentrations are apparent. For example,
 7 in New York City, we observe decreasing concentrations of certain metals in honey (Pb, Ti, V,
 8 362 Sb, Cd; i.e., metals often associated with anthropogenic activity) as a function of distance from
 9 central Manhattan: the highest concentrations are observed in honey from within the city (Fig.
 10 363 in New York City, we observe decreasing concentrations of certain metals in honey (Pb, Ti, V,
 11 Sb, Cd; i.e., metals often associated with anthropogenic activity) as a function of distance from
 12 364 central Manhattan: the highest concentrations are observed in honey from within the city (Fig.
 13 2, S1). On Kaua'i, honey on the south side of the island has uniquely elevated levels of certain
 14 365 metals (e.g., Pb, Cd, Ti, V, Zn, Mn) relative to honey from elsewhere on Kaua'i and from other
 15 Hawaiian Islands (Fig. 3, S2).
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 53 371 **Figure 3.** Metal concentration bivariate plots (Ti, V, Zn, Mn vs. Pb) for honey from four Hawaiian
 54 372 Islands: Hawai'i, Maui, O'ahu, and Kaua'i. All units are µg/g; note the logarithmic scale. For
 55 373 figure clarity, error bars are not included (errors are reported in Table S1).
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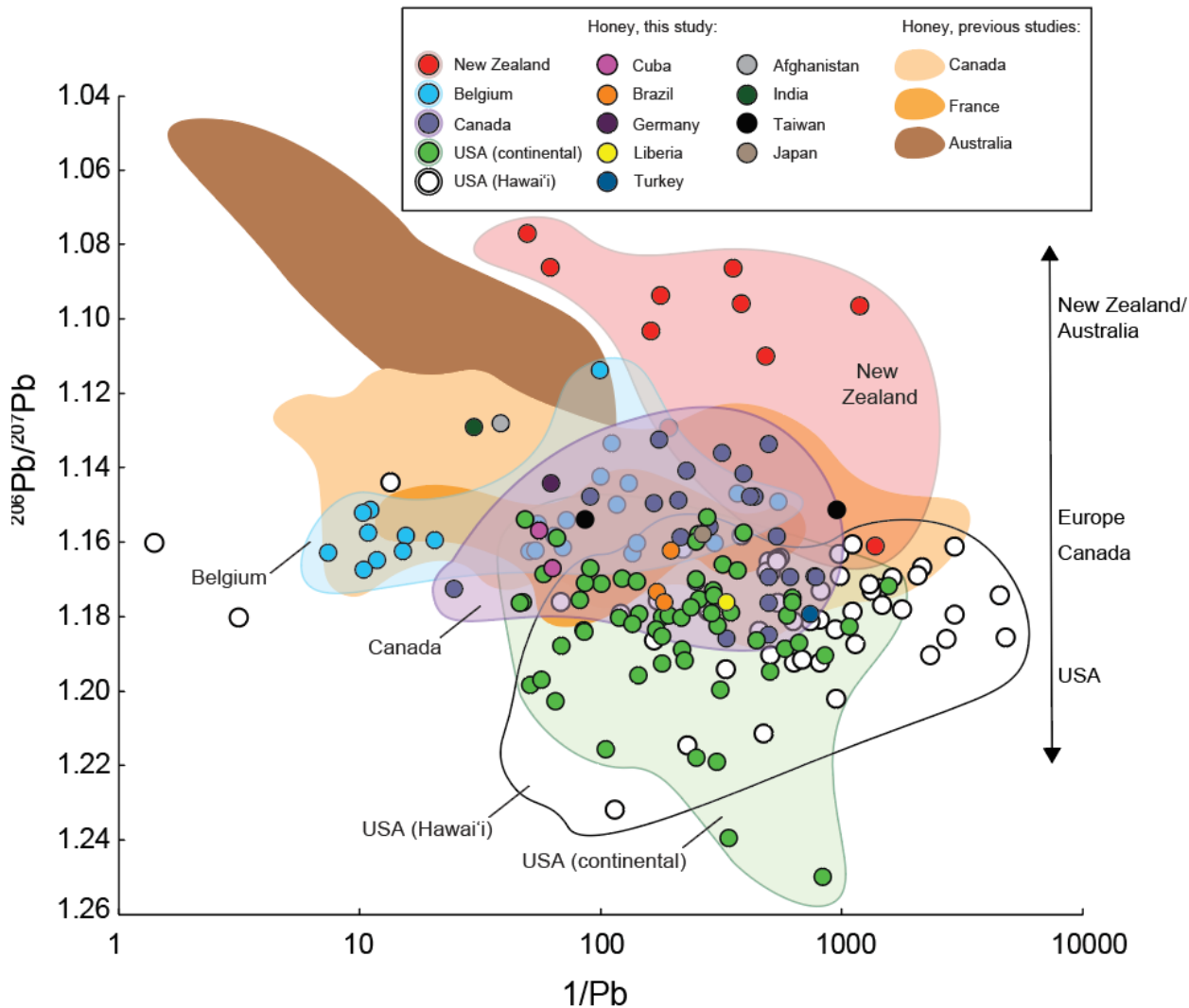


Figure 4. A plot of $1/\text{Pb}$ ($\text{g}/\mu\text{g}$) versus $^{206}\text{Pb}/^{207}\text{Pb}$ for all honey analyzed in this study. Honey from previous studies is included for context (as shaded fields): Broken Hill and Sydney, Australia (Zhou et al., 2018b); France (Smith et al., 2020); Vancouver, Canada (Smith et al., 2019; Smith and Weis, 2020). Regions from this study that include ≥ 10 samples are also shaded (Continental USA and Hawai'i, Canada, Belgium, New Zealand).

3.3 Pb isotopes

All honey samples analyzed in this study have a Pb isotopic range of 1.077 to 1.253 and 2.001-2.187 for $^{206}\text{Pb}/^{207}\text{Pb}$ and $^{208}\text{Pb}/^{206}\text{Pb}$, respectively (Table S1-S2). The Pb isotopic compositions of the honey vary based on their region of origin. In general, honey collected from N. America (especially from the eastern USA) is significantly more radiogenic (higher $^{206}\text{Pb}/^{207}\text{Pb}$, lower $^{208}\text{Pb}/^{206}\text{Pb}$) compared to European honey, and compared to honey from New Zealand, which

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has the least radiogenic isotopic compositions measured in this study ($p_{adj} < 0.05$ for $^{206}\text{Pb}/^{207}\text{Pb}$ and $^{208}\text{Pb}/^{206}\text{Pb}$, Fig. 1, 4). Based on geology and histories of anthropogenic Pb use (e.g., which Pb gasoline additives were used in various parts of the world), the isotopic results were as expected for all geographic regions given that they plot along the global Pb array, which is strongly governed by the twentieth century's leaded gasoline mixing line (Fig. 5). We observe one exception in this study: honey collected from the northeast part of Kaua'i deviates slightly from the Pb isotopic compositions of other Hawaiian honey (and the global Pb array), Hawaiian geology, and input from Eurasian dust (Fig. 6).

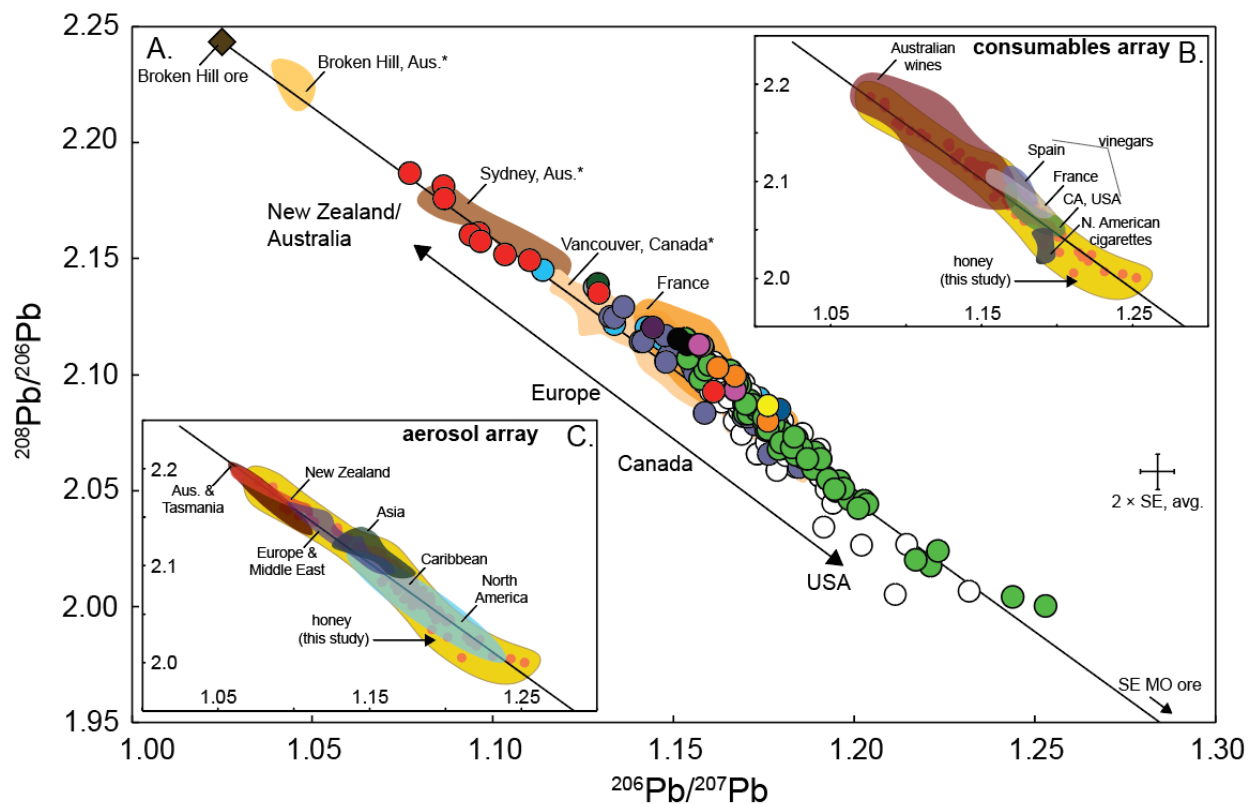


Figure 5. Panel A. Lead isotope plot of all honey analyzed in this study, with honey from previous studies included for context, as shaded fields (see Fig. 4 for symbol legend and caption for references). Data field labels denoted with an asterisk (*) include bees (and other hive products) in addition to honey. The Sydney field includes data for bees, honey, and wax from European honeybees (*Apis mellifera*) and an Australian native bee species (*Tetragonula carbonaria*) (Zhou et al., 2018b, 2018a). The line (featured in all panels) is a mixing line between the Broken Hill and Southeast Missouri (SE MO) ore endmembers; a representation of the isotopic compositions of leaded gasolines used globally (Gulson, 1984; Sangster et al., 2000). The average error bar ($2 \times$ standard error) is the mean error for all honey (except WI honey, for which error bars are smaller than the symbols). Panel B. Lead isotope plot of all honey data from this study (orange symbols in shaded yellow field), with various consumables from around

the globe for comparison: Australian wines (Kristensen et al., 2016), vinegars (apple cider, balsamic, wine, malt, cognac) (Ndung'u et al., 2011), and North American cigarettes (Guo et al., 2015). *Panel C.* Pb isotope plot of all honey data from this study (orange symbols in shaded yellow field), with the global aerosol array superimposed for comparison (Bollhöfer et al., 1999; Bollhöfer and Rosman, 2001, 2000).

The four galena samples from southwest Wisconsin have an isotopic range of 23.1996-23.8820, 16.1133-16.1797, and 42.750-43.281 for $^{206}\text{Pb}/^{204}\text{Pb}$, $^{207}\text{Pb}/^{204}\text{Pb}$, and $^{208}\text{Pb}/^{204}\text{Pb}$, respectively (Table S2). These results are within the Pb isotopic range reported for other MVT Pb deposits from the Upper Mississippi Valley District and are in good agreement with other ores from southwest Wisconsin, which have notably radiogenic compositions (Fig. 7, 8) (Millen et al., 1995).

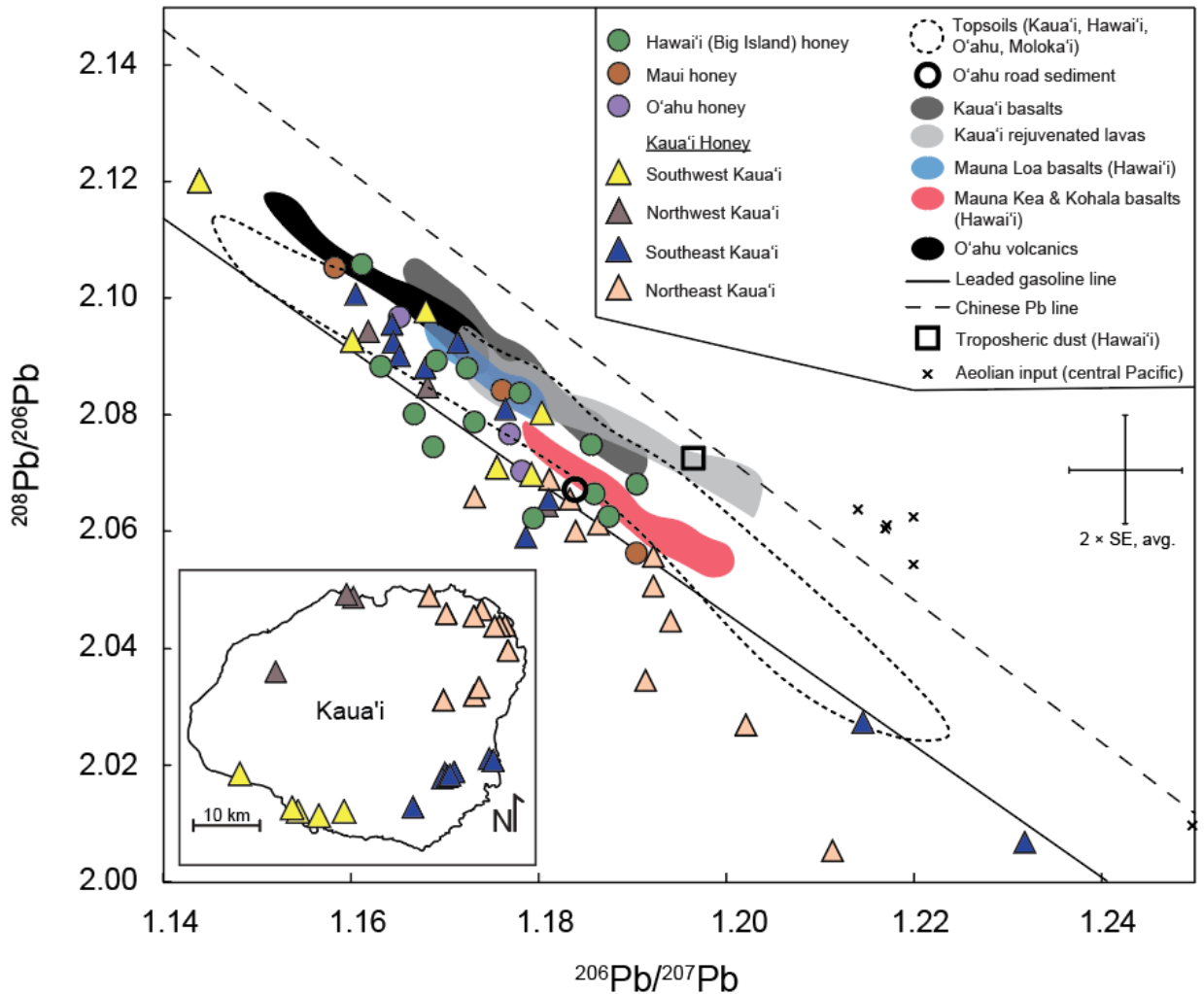


Figure 6. Lead isotope plot of honey collected from four Hawaiian Islands: Hawai'i, Maui, O'ahu,

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and Kaua'i. For context, Pb isotopic compositions of local geology is included: basalts from Mauna Loa, Mauna Kea, and Kohala volcanoes (Abouchami et al., 2000; Hanano et al., 2010; Weis et al., 2011), Kaua'i basalts and O'ahu volcanics (Wai'anae and Ko'olau) (Williamson et al., 2019). Topsoils (dotted field) are compiled from Monastra et al., (2004) and Spencer et al., (1995) and O'ahu road sediments are from Sutherland et al. (2003). The gasoline mixing line is the same as described in Figure 5. The Chinese Pb line is defined by ores, coals, and fuel emissions from China (Bi et al., 2017; Cheng and Hu, 2010). Other aeolian inputs include samples collected from central Pacific Ocean sites surrounding the Hawaiian Islands (Jones et al., 2000) and tropospheric silicate dust collected on Hawai'i (Spencer et al., 1995). Inset map: Kaua'i honey sampling locations.

4. DISCUSSION

4.1. Eastern North America and the Great Lakes region

4.1.1. New York Metro Area

Cities tend to store more metals (especially Pb, as it lingers in the environment), since they host myriad industrial centers, roadways, ports, etc., in a concentrated area; the size and age of the city compound this effect (Chambers et al., 2016; Laidlaw and Filippelli, 2008; Mielke et al., 2011). Lead concentrations in the NY Metro honey exhibit a similar distribution of Pb in topsoil collected throughout NYC: higher concentrations in areas of high traffic, industrial activity, and in the oldest parts of the city, with the highest Pb concentrations measured in soils from Manhattan, Brooklyn, and Queens (Cheng et al., 2015; Li et al., 2018; Paltseva, 2019). The effect of nearby land use on other metals in honey in these regions is evident: for example, elevated Sb (vehicle brake wear) (Von Uexküll et al., 2005), Zn (galvanized structures, rubber tire crumbs) (Huber et al., 2015), and V (fuel oil combustion) (Zhao et al., 2013) (Fig. 2, S1). In a study of vegetables (root, leafy, and juicy varieties) from urban gardens across NYC, McBride et al. (2014) determined that Pb content is mostly a result of physical contamination of resuspended soil particulates and aerosol deposition, both of which are proven pathways for environmental particulates to make their way into honey (Negri et al., 2015; Pellicchia and Negri, 2018; Zhou et al., 2018b). Thus, typical city activities that contribute to emissions and particulate resuspension, such as heavy stop-and-go traffic, shipping ports, demolition, and construction, are all contributing to elevated levels of metals in NY Metro honey, relative to levels measured in honey collected further from the city center. Similar metal concentration gradients, as a function of distance from the city center, were observed in honey from Metro Vancouver,

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456 Canada, and Paris, France (Smith et al., 2020, 2019; Smith and Weis, 2020), with some logical
457 differences (e.g., there is minimal spatial variation in V concentration in the non-port city of
458 Paris, compared to New York and Vancouver that host major shipping ports) (Fig. 2, insets).

459 The Pb isotopic range of the NY Metro honey ($^{206}\text{Pb}/^{207}\text{Pb}$: 1.154-1.195) is consistent
460 with ranges previously reported for topsoil, dust, and paints throughout the NY Metro Area.
461 Some honey from this study is slightly (not significantly) less radiogenic than that of outdoor
462 dust collected throughout the city from the tops of pedestrian crossing signals at major
463 intersections ($^{206}\text{Pb}/^{207}\text{Pb}$: 1.17-1.23) (Caravanos et al., 2006). Our honey range overlaps with
464 that of house paints ($^{206}\text{Pb}/^{207}\text{Pb}$: 1.141-1.186) and indoor dust samples ($^{206}\text{Pb}/^{207}\text{Pb}$: 1.166-
465 1.192) collected in Hillsdale, NJ (part of the NY Metro Area) (Jaeger et al., 1998). The NY Metro
466 honey data also falls within the Pb isotopic ranges measured in indoor house dust ($^{206}\text{Pb}/^{207}\text{Pb}$:
467 1.134-1.227), interior paints (1.092-1.267), garden soils (1.612-1.200), and street dust (1.155-
468 1.199) collected in Jersey City, NJ (immediately west of Lower Manhattan, across the Hudson
469 River, Adgate et al., 1998), and topsoil from Staten Island ($^{206}\text{Pb}/^{207}\text{Pb}$: 1.144-1.196, Pribil et al.,
470 2014).

471 While the Pb isotopic compositions of NY Metro honey fall within the expected isotopic
472 range of the region (Fig. 7), the results do not reveal a predictable geospatial gradient or trend
473 at city-scale, in contrast to previous studies in Metro Vancouver (Smith et al., 2019; Smith and
474 Weis, 2020) and Sydney, Australia (Zhou et al., 2018b). This is likely due to NYC being much
475 larger and older than both Metro Vancouver and Sydney, meaning modern and historical land
476 use, zoning, and industry are comparatively more extensive and convoluted in NYC. Indeed, a
477 similar honey survey, recently undertaken in an even older city (Paris, France), showed a
478 completely predictable Pb concentration gradient (matching the footprint of the fallout from
479 the 2019 fire at Notre-Dame cathedral), while the Pb isotopic compositions of honey exhibited
480 no geospatial trends within the city due to millennia-scale history of Pb use in Paris (Smith et
481 al., 2020). The Paris study and the NY metro honey results imply that the utility of km-scale
482 geospatial resolution of Pb isotopic compositions in honey is not only a function of land use and
483 recent historical development but of the comprehensive geochemical ‘baseline’ of the region,
484 i.e., all natural and anthropogenic Pb inputs, to date (Matschullat et al., 2000).

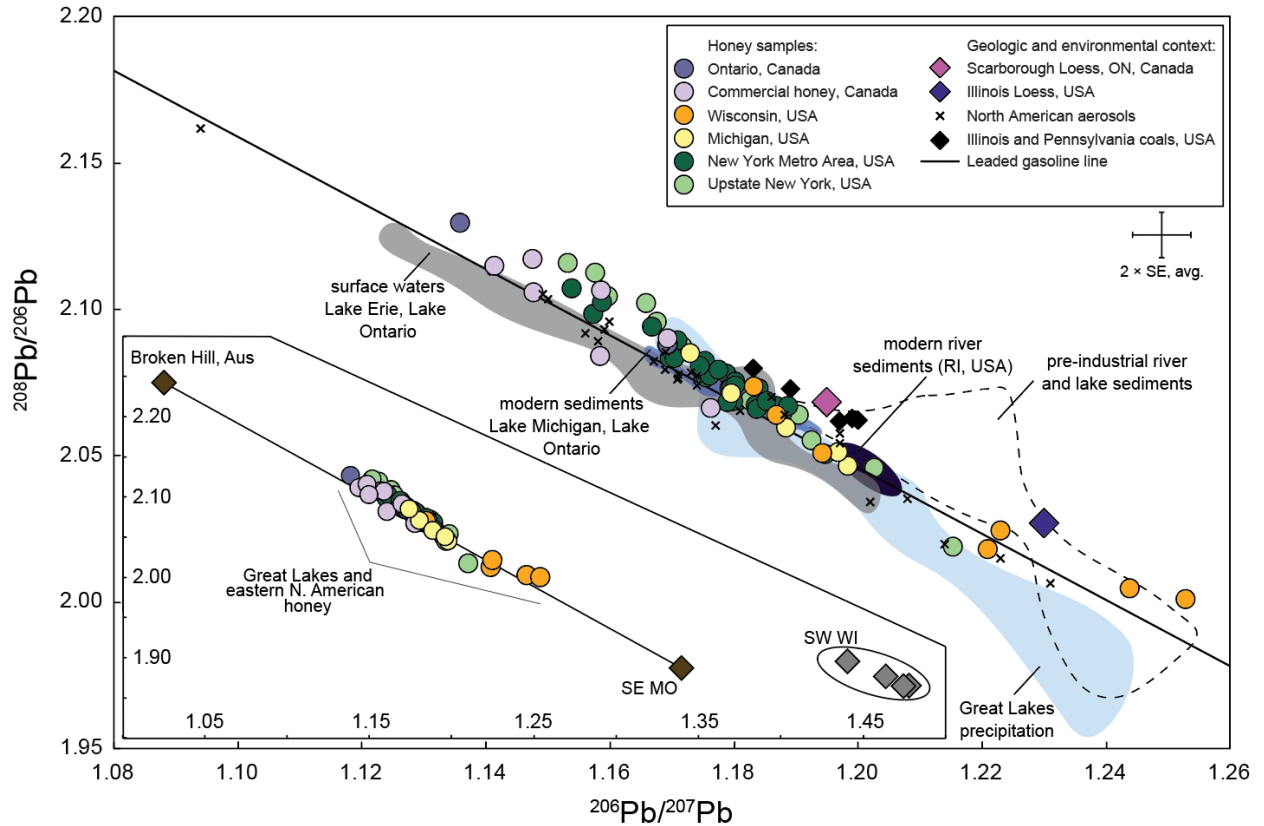


Figure 7. Honey from the Great Lakes region/eastern North America. For commercial honeys from Canada, we only included honey brands that claimed to be of Canadian origin (more specific origin information is unknown). For other honey, if specific origin location is known, it is listed in Table S1. Main plot: Selected local and regional context is provided from the literature, including North American aerosols (Bollhöfer and Rosman, 2001), loess (Biscaye et al., 1997), coals from Illinois and Pennsylvania (Chow and Earl, 1972), Great Lakes region precipitation (Sherman et al., 2015), Great Lakes (Erie and Ontario) surface waters (Flegal et al., 1989), modern (1990s and later) lake sediments from inland lakes in Michigan and Ontario (Cheyne et al., 2018), and modern sediments (1980s and later) from the Pettaquamscutt River (Lima et al., 2005). The pre-industrial field (dashed line) consists of lake and river sediments from the 1700s (Cheyne et al., 2018; Graney et al., 1995; Lima et al., 2005). The gasoline mixing line is the same as described in Figure 5. Average error bars ($2 \times$ standard error) are for all honey (except the WI honey, which has error bars smaller than the symbols). Inset: A simplified version of the main panel, zoomed out to show endmembers of the gasoline mixing line and other Pb ores. 'SW WI' is galena from a MVT ore deposit in southwest Wisconsin (Lafayette Co.).

4.1.2. Pb isotopes in USA versus Canadian honey

Lead isotopic compositions of honey from eastern USA and Canada are comparable to Pb isotopic ranges reported in other environmental matrices in that region: precipitation, snow, and modern lake sediments and surface waters, and all honey from this region plots within

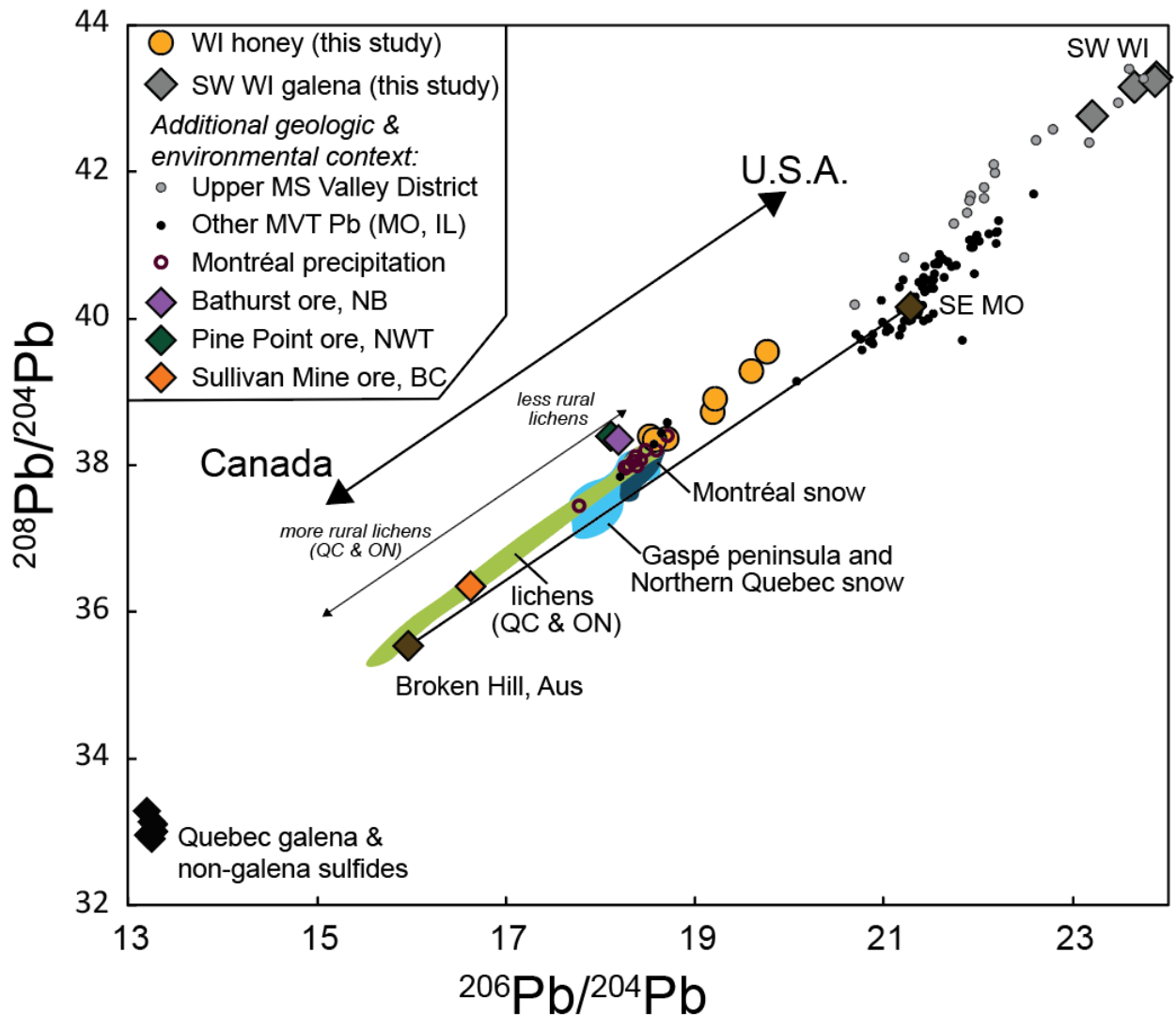
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error of the leaded gasoline mixing line (Fig. 7 and references therein). Honey samples from the USA have Pb isotopic compositions that are more radiogenic than those from Canada ($p_{adj.} < 0.001$ for $^{206}\text{Pb}/^{207}\text{Pb}$ and $^{208}\text{Pb}/^{206}\text{Pb}$). This can be explained by differing industrial histories in these regions and different sources of raw materials for Pb smelting, manufacturing, and recycling in each country. For instance, both countries used some amount of leaded gasoline mixed with tetraethyl Pb (TEL) produced from Southeast Missouri (SE MO) Pb ore, but Canada used only a small percentage in comparison to the USA and sourced the majority of their TEL from British Columbia and New Brunswick (Canada imported < 1 % of Pb ore used for TEL), while the USA supplemented their SE MO Pb with ores from Australia, Mexico, and Peru (Sturges and Barrie, 1987). The USA also had considerably more industrial activity, consuming more Pb and other raw materials than Canada (from ~1750 to the present), resulting in present-day Pb enrichment factors (EF_{Pb}) in lake sediments of ~100 in the Great Lakes region (USA side) compared to EF_{Pb} of ~10 in lake sediments from eastern Canada (Marx et al., 2016).

In the 1980s, when leaded gasoline was still in widespread use, dissolved Pb samples from Lakes Erie and Ontario had more radiogenic compositions when the wind was blowing from the south/USA and were conversely less radiogenic when the wind was blowing from the north/Canada (Flegal et al., 1989). Today, it seems that ongoing sulfide mining and Pb refining in the Mississippi Valley and Quebec (Abitibi Greenstone Belt) continue to ensure that honey from the eastern USA and Canada have differing Pb isotopic compositions, especially since the major metropolitan areas of the USA are downwind of key industrial and mining regions of the mid- and eastern USA, due to prevailing winds (Klink, 1999). This is further clarified when using the ^{204}Pb isotope to examine this disparity; honey from WI plots closer to the USA/MVT Pb ores, as expected based on context from USA and Canadian Pb ores and environmental samples (precipitation, snow, lichens) from both rural and urban areas in eastern Canada (Fig. 8 and references therein).

Earlier studies that focused on city-scale nuance of Pb distribution in honey demonstrated that honey can resolve input from point sources or acute Pb pollution events (Smith et al., 2019; Smith and Weis, 2020; Zhou et al., 2018b), but this localized variation is lost at the regional/sub-continental scale, which better reflects regional-scale processes, i.e.,

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 4 535 different Pb ore endmembers used in the eastern USA versus eastern Canada in this case. Metal
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 6 536 concentrations and Pb isotopic compositions of sediment cores from southern Ontario and
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 8 537 northern Michigan recorded the history of local and regional metal input, which, at certain
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 10 538 points in time, overwhelmed that of the continental, global leaded gasoline trends of the
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 12 539 twentieth century (Cheyne et al., 2018). Presumably, if honey had been collected at a higher
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 14 540 resolution near Great Lakes industrial centers throughout the last century (e.g., Toronto,
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 16 541 Chicago, Milwaukee, Toledo), it would reveal the local nuances of industrial pollution at each
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 18 542 site.



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 56 544 **Figure 8.** Lead isotope plot featuring honey from south-central WI. The honey plots toward the
 57 545 USA end of the Pb mixing line defined by differing emissions in eastern Canada vs. eastern USA,
 58 546 e.g., after Simonetti et al. (2000a). Reference reservoirs include SW WI galena (this study),
 59 547 Upper Mississippi (MS) Valley District MVT ores (Millen et al., 1995), other USA MVT Pb ores

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(Goldhaber et al., 1995), Canadian (Quebec) Pb ores (galena and non-galena sulfides) from the Atibiti Pb sulfide belt (Franklin et al., 1983), and other economically significant Pb ores (mined throughout the past century) in Canada: Pb from the Sullivan Mine (Kimberley, BC), the Pine Point Mine (Great Slave Lake, NWT), and the Bathurst Mining District (northeast NB) (compiled by Sangster et al., 2000). Environmental context includes Montréal precipitation data (Simonetti et al., 2000b), and snowpack and lichen data from Montréal, moderately remote areas (e.g., the Gaspé Peninsula/St. Lawrence Valley), and very remote areas (i.e., along a north-south transect between southern ON/QC and Hudson Bay) (Carignan and Gariépy, 1995; Simonetti et al., 2000a). Error bars ($2 \times$ standard error) for the WI honey and southwest WI galena data are smaller than the symbols.

4.2 Kaua’i, Hawaiian Islands

Major sources of Pb on the Hawaiian Islands are lithogenic (Hawaiian basalts), and legacy anthropogenic (leaded gasoline, leaded paint, discontinued/banned pesticides, e.g., lead arsenate). There is also some aeolian dust input from northeast Asia, confirmed by its continental-derived $^{87}\text{Sr}/^{86}\text{Sr}$ ratios (too high to be derived from the basaltic, Hawaiian geology) measured in Hawaiian topsoil (Kurtz et al., 2001; Spencer et al., 1995). Lead source apportionment calculations for road-deposited sediments from O’ahu indicate inputs from gasoline additives of 31-81 % in sediment collected after 1968, the period when anthropogenic Pb input was more easily identified, since TEL content of USA gasoline shifted toward increased usage of the very radiogenic SE MO ores (Sutherland et al., 2003). Compared to honey from the other Hawaiian Islands, Kaua’i honey exhibits the largest range in Pb isotopes and the most deviation from lithogenic Pb isotopic compositions, including basalts from Kaua’i and other Hawaiian islands (Hanano et al., 2010; Williamson et al., 2019) and loess transported from the Chinese Loess Plateau and deposited in the central Pacific (Jones et al., 2000). The largest Pb isotopic ranges are observed in honey from the eastern side of the island, where most of Kaua’i’s small population (~72,200 people total) reside, so the observed isotopic compositions are likely a result of current and legacy anthropogenic activity (e.g., backyard scrapping of old vehicles, old paint, legacy gasoline, agricultural activity), especially since any deviations from the global gasoline mixing line (i.e., honey from northeast Kaua’i) do not plot in the direction of Kaua’i’s natural/geogenic Pb isotopic compositions. For comparison, commercially available honey from Hawai’i Is. has Pb isotopic compositions that plot between natural and anthropogenic Pb isotopic compositions (Fig. 6).

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581 Large-scale agricultural activity is likely contributing to the anomalously high metal
582 concentrations observed in Kaua'i honey compared to honey from O'ahu, Maui, and Hawai'i
583 (Fig. 3, S2). The south side of Kaua'i is mostly in seed crop and coffee production (Spengler et
584 al., 2019). A previous study of Kaua'i honey established that concentrations of glyphosate (an
585 organophosphate herbicide marketed as Roundup™ by Monsanto/Bayer AG) measured in
586 honey correlated well with land use: hives near large-scale agriculture operations, roadsides, or
587 golf courses produced honey with higher concentrations of glyphosate (Berg et al., 2018). The
588 regions where Berg et al. (2018) observed the highest glyphosate concentrations in honey are
589 in the south/southwest of the island, where we observe some of the highest metal
590 concentrations (in particular, Pb, Ti, V, Cr, Fe, Ga, Zr, Ba, Fig. 3, S2) observed in honey, not only
591 from the Hawaiian Islands, but in any honey analyzed for this study (Figs. 1, 4, Table S1).
592 Pesticide and fertilizer use as a source for some of these metals should not be discounted,
593 including Cd, Co, Cu, Ni, Pb, and Zn (Gimeno-García et al., 1996). Defarge et al. (2018) measured
594 As, Co, Cr, Ni, and Pb (ranging from 10s to 100s ng/g) as formulant ingredients in glyphosate-
595 based herbicides and other commercially available herbicides, fungicides, and insecticides.
596 Fertilizer, including phosphate fertilizers, sewage sludges, manures, and fly ash (compiled by
597 Kabata-Pendias, 2010 and U.S. EPA, 1999), also introduces metals of foreign origin into
598 agricultural fields. Additionally, three honey samples collected in southern Kaua'i were from
599 feral hives, one within an abandoned building, all three within 0.5 km of abandoned farming
600 vehicles and equipment.

601 Another important contributing factor for metal transport is climate. Kaua'i's central,
602 high-elevation regions experience large amounts of rainfall; up to 1140 cm per year on Mt.
603 Waialeale, among the highest amounts measured anywhere on Earth (US National Weather
604 Service, 2020), which leads to immense amounts of erosion and washout into the valleys
605 (Ferrier et al., 2013) even on the comparatively 'dry' south/southwest side of the island. The
606 Waimea River, which empties into Waimea Bay (mean daily discharge rate of 1.1 m³/s, 63-years
607 of data, USGS, 2020), is fed by a high-elevation swamp that receives significant amounts of
608 rainfall. Coupled with a humid, tropical climate, extreme precipitation eventually reduces soil
609 pH and mobilizes heavy metals from soils (Spencer et al., 1995). On O'ahu, others (Andrews and

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610 Sutherland, 2004; Heinen De Carlo and Anthony, 2002) observed that metals of both
611 anthropogenic and lithogenic origin were transported downstream where they gradually
612 accumulated in the upper soil horizons and sediment layers relative to background levels, so a
613 similar transport mechanism is likely on Kaua'i (which has undergone a similar geologic and soil-
614 forming history) (e.g., Orazio et al., 2007).

615 A new, modern study on the geochemistry of Kaua'i topsoil, C-horizon soils, sediments
616 from Waimea Bay (downstream of the Waimea Valley), e.g., after Spencer et al. (1995), and
617 honey would be worthwhile. In particular, sediments would offer the temporal context
618 necessary to reveal whether the downstream sediments contained lower metal contents prior
619 to large-scale farming efforts in the past decades. The ongoing effects of metal and pesticide
620 transport in this region and assessment of the polluting role (if any) that other unique land use
621 features on Kaua'i (e.g., the US Navy Pacific Missile Range Facility, the Port Allen solar voltaic
622 power plant, and legacy diesel-fed sugar mills) may have on the region should remain the
623 subject of ongoing and future studies. Higher-resolution honey sampling on the other Hawaiian
624 Islands would provide a useful comparison between the Pb isotopic compositions and metal
625 concentrations observed on Kaua'i versus those in larger Hawaiian urban centers (e.g.
626 Honolulu, O'ahu).

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628 **4.3 Honey and the global Pb array**

629 The Pb isotopic compositions of honey fit predictably on the global Pb isotope array based on
630 the geographic origin of the honey. For example, honey samples from New Zealand have Pb
631 isotopic compositions that are distinctly less radiogenic compared to honey from western
632 Europe, Canada, Hawai'i, and the continental USA, while New Zealand compositions are
633 comparable to other samples from Oceania: bees, honey, beeswax, topsoil, and dust from
634 Sydney, Australia, red wines from South Australia (Kristensen et al., 2016), and aerosols from
635 New Zealand, Tasmania, and Australia (Bollhöfer and Rosman, 2000). The least radiogenic Pb
636 isotopic compositions reported for honey, bees, and other hive products were collected in the
637 vicinity of the Broken Hill Mine, Australia (Zhou et al., 2018b), an area noted for providing one
638 of the two Pb ores used to produce the majority of the world's TEL (the other being SE MO ore,

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4 639 at the other end of the Pb isotopic composition spectrum for TEL used globally) (Fig. 5a).

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6 640 Western European honey from this study (Belgium and Germany) have similar Pb
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8 641 isotopic compositions to previously reported values for honey from France (from Paris and the
9
10 642 Auvergne-Rhône-Alpes regions, Smith et al. 2020) and have similar Pb isotopic compositions to
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12 643 wines and vinegars from western Europe (France, Spain, Italy) (Fig. 5b, and references therein).
13
14 644 Honey and other products from the USA tend to have the most radiogenic Pb isotopic
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16 645 compositions, with honey from Canada falling between the USA and western European ranges.
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18 646 This is likely due to consumption of TEL from differing sources throughout the twentieth
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20 647 century. There are few continental-scale isoscapes available for Pb, but the honey results match
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22 648 isoscapes mapped using other environmental proxies: the Pb isotopic compositions for topsoil
23
24 649 and tooth enamel from the USA are generally more radiogenic than those observed in western
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26 650 European isoscapes constructed using similar proxies (Keller et al., 2016; Reimann et al., 2012,
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28 651 2011).

29
30 652 Unsurprisingly, the limited honey samples from regions other than North America,
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32 653 Western Europe, and New Zealand (Table S1, Fig. 5a) also plot along the global Pb array. These
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34 654 are the first Pb isotopic compositions reported for honey from Afghanistan, Brazil, Cuba,
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36 655 Germany, Liberia, Taiwan, and Turkey and may serve as a starting point for future studies in
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38 656 those regions. Confirming the geographic authenticity of these honey samples is beyond the
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40 657 scope of this study, but their Pb isotopic compositions fit well with existing Pb aerosol data
41
42 658 from the same region/country (Fig. 5). For example, the two Cuban honey blends and three
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44 659 Brazilian honey samples analyzed in this study have Pb isotopic compositions in agreement with
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46 660 ranges reported for aerosols in Cuba and Brazil (collected at the end of the twentieth century)
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48 661 (Bollhöfer and Rosman, 2000).

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50 51 663 **4.4 Geochemical implications of sampling resolution**

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53 664 A city-scale survey of metals in honey provides insight into factors that influence metal
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55 665 distribution locally, and the corresponding Pb isotopic compositions can provide insight into
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57 666 metal sources. These factors may include local topography, land use, municipal-level
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59 667 environmental policy and regulation, construction, and acute pollution events (Smith and Weis,

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668 2020). For example, the metal concentration data for honey from the New York Metro Area and
669 Kaua'i reveal interesting local patterns of metal distribution. The metal concentrations reported
670 in this study were selected for their utility in geochemical investigations; these metals are most
671 useful on a local scale when combined with a high sampling resolution to differentiate between
672 anthropogenic input versus local background, i.e., best for sourcing metals as opposed to
673 tracing the geologic source of the honey. Thus, these metals are not directly useful for
674 comparison between global regions since there is too much variation in metal mobility as a
675 result of local factors such as plant diversity and soil properties (pH, porosity, organic material,
676 and clay content) (Kabata-Pendias, 2010, 2004). The suite of metals reported in this study may
677 nonetheless supplement future honey surveys that include additional analytes that do assist in
678 resolving the region of origin for honey or to confirm honey authenticity and/or botanical
679 origin. These include certain macronutrients (K, P) (Chen et al., 2014; Zhou et al., 2018c) and
680 lanthanide series metals (Pellerano et al., 2012).

681

682 *4.4.1 Geochemical 'baseline' and future of Pb isotope use in honey*

683 Honey meets the requirements suggested by Darnley et al. (1995) for adding viable data to
684 a global geochemical database. These requirements include being a commonly available
685 sampling matrix where adequate amounts can be stored indefinitely for future reference. Lead
686 isotopic compositions of wine samples from France (Epova et al., 2020) and Australia
687 (Kristensen et al., 2016) with vintages spanning the twentieth-century exhibit a temporal trend,
688 reflecting the introduction and subsequent phase-out of leaded gasoline in their respective
689 geographical regions. Presumably, if honey archives existed throughout the past two centuries,
690 they would also reflect the key moments of Pb use throughout recent history. Like wine, honey
691 is easily stored for very long periods of time, if not indefinitely. Given that the most recent and
692 advanced spectroscopic techniques (e.g., NMR, FT-IR) are becoming common-place for rapid-
693 throughput authentication efforts (Wu et al., 2017), perhaps archiving of honey will become
694 standard practice, creating a chemical repository for various environmental investigations, i.e.,
695 as a record of past environmental conditions and a continuous record of baseline compositions.

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696 Establishing a baseline (natural/background input + anthropogenic input to-date) for a
697 given chemical species is important in geochemical surveys (Matschullat et al., 2000; Smith et
698 al., 2018). Since these data reflect the current baseline for Pb isotopic compositions in honey,
699 they represent a useful tool for future comparison as Pb use shifts. However, confidently
700 identifying the contributions of natural *versus* anthropogenic Pb in consumer products in the
701 modern, global economic climate is becoming increasingly difficult for three key reasons. First,
702 recent (since the mid-1800s) anthropogenic input of Pb into the environment is large (1 to 2
703 orders of magnitude larger) compared to natural inputs (Han et al., 2002; Nriagu, 1989; Weiss
704 et al., 1999). Second, all anthropogenic sources were originally derived from natural sources –
705 this is obvious but depending on the scale of a given Pb investigation (local/city-scale,
706 regional/continental, global), this concept can conflate Pb mixing models unless possible Pb
707 inputs are identified and quantified at each scale (Flegal and Odigie, 2020). Third, ores and raw
708 materials are shipped globally for refining and manufacturing, then manufactured goods are
709 shipped elsewhere, and post-consumer materials (e.g., Pb batteries) may be shipped
710 internationally yet again for recycling. It is important to note that even raw materials from the
711 same ore deposit can exhibit natural Pb isotopic variability (e.g., Sangster et al., 2000). The food
712 industry has also been globalized, in particular for specialty, non-perishable items like honey
713 where claims of ‘terroir’ or source environment can be economically valuable, e.g., the manuka
714 honey industry of New Zealand (Carter et al., 2016; Rogers et al., 2014). In general, the honey
715 market remains non-transparent. For example, the distributor of commercially-available honey
716 purchased in North Carolina, USA, for this study stated that the product was a blend of
717 American honey and imported honey but would not disclose any further details about mixing
718 ratios or the geographic source locations of the honey (Pb isotopic composition of this honey is
719 consistent with other North American honey, however).

720 For honey and other biomonitoring applications, we suggest using Pb isotopes in
721 conjunction with other tools to strengthen sourcing studies. For example, analysis of Sr or Nd
722 isotopes in honey may help distinguish geogenic inputs, e.g., dust and endmember loesses
723 (Grousset and Biscaye, 2005), from anthropogenic inputs (Baroni et al., 2015; Voerkelius et al.,
724 2010). We also suggest more extensive use of the ²⁰⁴Pb isotope in environmental Pb surveys

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4 725 (e.g., Fig. 8) to verify that Pb sourcing interpretations are not oversimplified (e.g., Ellam, 2010).
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6 726 If the goal is to trace the origin of the honey, rather than the source of the Pb and other metals,
7
8 727 there is no doubt that a global geochemical database for honey (in which sampling origin is
9
10 728 known and knowledge exists regarding anthropogenic and geogenic sources of metals in honey)
11
12 729 will supplement and strengthen the existing statistical approaches that use carbon isotopes and
13
14 730 trace metal data (Zhou et al., 2018c) and the higher-throughput spectroscopy methods of
15
16 731 honey sourcing and authentication, e.g., various infrared and nuclear magnetic resonance
17
18 732 techniques (Wu et al., 2017).
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21 734 **5. CONCLUSIONS**

23 735 This study presents Pb isotopic and metal concentration data for honey at three different
24
25 736 scales: local, with high-resolution sampling (NY Metro Area and Kaua'i), regional (Eastern North
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27 737 America), and global. Local-scale honey data gives insight into local processes affecting metal
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29 738 distribution and aids in identifying point and/or diffuse sources of metals. Metal concentrations
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31 739 in honey indicate potential effects of large-scale farming in and around Waimea on Kaua'i and
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33 740 typical, city-type distribution gradients of metals associated with human activity (e.g., Pb, Sb, V,
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35 741 Zn) throughout New York City. On these small geographic scales, there is some variation in Pb
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37 742 isotopic composition, but their range is largely contained within the isotopic constraints defined
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39 743 by local sources of Pb, whether natural or anthropogenic. In larger-scale geochemical surveys
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41 744 using honey, metal concentrations are less useful for mapping distributions and identifying
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43 745 point sources and the Pb isotopic compositions reflect regional and global processes related to
44
45 746 Pb. A regional survey of Pb isotopic compositions in honey reflects the industry, urbanization,
46
47 747 and geology of eastern North America, revealing differences in modern Pb emissions in the
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49 748 eastern USA and eastern Canada. Globally, Pb isotopes in honey roughly reflect the continental
50
51 749 origin of the honey: samples from the USA are most radiogenic and honey from New Zealand is
52
53 750 least radiogenic. The global trend predictably matches the modern global Pb array, as defined
54
55 751 by aerosols, varying regional geologic background, and nearly a century of leaded gasoline use
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57 752 around the world. This study provides a sound starting point for building current baseline data
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59 753 for honey as a biomonitor for metals in the regions discussed here and provides a global

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4 754 framework for Pb isotopic compositions in honey, which is useful context for more localized
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6 755 studies. This study also demonstrates the utility of a community science effort for honey
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8 756 collection.
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10 757 **SUPPLEMENTARY MATERIAL**

11 758 -Tables S1-S3

12 759 -Figures S1-S2

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29 776 **AUTHOR CONTRIBUTIONS**

30 777 **K. Smith:** investigation, formal analysis, validation, writing-original draft, writing-review and
31 778 editing, visualization, resources **D. Weis:** conceptualization, funding acquisition, writing-review
32 779 and editing, supervision **S. Scott:** resources, investigation, validation, writing-review and editing
33 780 **C. Berg, Y. Segal, P. Claeys:** resources, writing-review and editing.
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35 782 **CONFLICT OF INTEREST**

36 783 The authors declare no conflicts of interest relevant to this study.
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Regional and global perspectives of honey as a record of lead in the environment

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CONFLICT OF INTEREST STATEMENT

The authors declare no conflicts of interest relevant to this study.

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: